Quantitative sampling and analysis of trace elements in ambient air: impactor characterization and Synchrotron-XRF mass calibration

A. Richard\textsuperscript{1}, N. Bukowiecki\textsuperscript{1}, P. Lienemann\textsuperscript{2}, M. Furger\textsuperscript{1}, B. Weideli\textsuperscript{2}, M. Fierz\textsuperscript{3}, M. C. Minguillón\textsuperscript{1,4}, R. Figi\textsuperscript{5}, U. Flechsig\textsuperscript{6}, K. Appel\textsuperscript{7}, A. S. H. Prévôt\textsuperscript{1}, and U. Baltensperger\textsuperscript{1}

\textsuperscript{1}Laboratory of Atmospheric Chemistry, Paul Scherrer Institut, Villigen, Switzerland
\textsuperscript{2}School of Life Sciences and Facility Management, Wädenswil, Switzerland
\textsuperscript{3}University of Applied Sciences Northwestern Switzerland, Windisch, Switzerland
\textsuperscript{4}Institute for Environmental Assessment and Water Research (IDAEA), CSIC, Barcelona, Spain
\textsuperscript{5}Empa, Swiss Federal Laboratories for Materials Testing and Research, Dübendorf, Switzerland
6 Swiss Light Source, Paul Scherrer Institut, Villigen, Switzerland
7 Hamburger Synchrotronstrahlungslabor at Deutsches Elektronen-Synchrotron DESY, Research Centre of the Helmholtz Association, Hamburg, Germany

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Correspondence to: A. Richard (agnes.richard@psi.ch)

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Abstract

Identification of trace elements in ambient air can add substantial information to pollution source apportionment studies, although they do not contribute significantly to emissions in terms of mass. A method for quantitative size and time-resolved trace element evaluation in ambient aerosols with a rotating drum impactor and synchrotron radiation based X-ray fluorescence is presented. The impactor collection efficiency curves and size segregation characteristics were investigated in an experiment with oil and salt particles. Cutoff diameters were determined through the ratio of size distributions measured with two particlesizers. Furthermore, an external calibration technique to empirically link fluorescence intensities to ambient concentrations was developed. Solutions of elemental standards were applied with an ink-jet printer on thin films and area concentrations were subsequently evaluated with external wet chemical methods. These customized and reusable reference standards enable quantification of different data sets analyzed under varying experimental conditions.

1 Introduction

Information on temporal variation and size segregation of trace elements in ambient air greatly facilitates the identification of pollution sources. Particulate matter (PM) emissions caused by traffic exhaust and combustion processes are predominant in the fine size range (particle diameter<1 µm). In contrast, mechanically produced particles and mineral or resuspended road dust particles are mainly found in the coarse size fraction (particle diameter>1 µm). Most emissions such as traffic and industry related emissions or atmospheric dilution processes vary within a few hours. Conventional 24-h filter measurements do not detect these rapid changes due to the missing time resolution. Therefore, cascade impactors with lower measurement intervals (in the order of hours) as developed by Lundgren (1967) are very valuable in combining size segregation with high time resolution. A variety of impactors with different size cuts, number
of impaction stages, number of nozzles and substrates for various analysis methods exist (Hering, 1979; Raabe et al., 1988; Berner et al., 1980; D’Alessandro et al., 2003; Marple et al., 1991; Marple, 2004). Cahill et al. (1985) and Cliff et al. (2003) designed a three-stage rotating drum impactor. Bukowiecki et al. (2009) presented a modified design to obtain particle cutoff sizes of 2.5 µm, 1 µm, and approximately 0.1 µm. They included a first characterization of the impactor by determining the collection efficiency curves. The highly time resolved measurements of trace elements in ambient air result in low amounts of sample material in the range of a few µg per analyzed area. This demands a highly sensitive detection method such as synchrotron radiation based X-ray fluorescence spectrometry (SR-XRF), which provides a high sensitivity on small analysis areas. Bukowiecki et al. (2008) established an automated procedure to analyze many spectra in a reasonable time.

Since deviations relative to the desired size cuts would result in an incorrect size attribution of the particulate matter, knowledge about the size segregation characteristics of the impactor is crucial for the data quality. This paper describes the determination of the inverse collection efficiency curves for the three size ranges by the use of an artificial aerosol generator. The rotating drum impactor is introduced in Sect. 2.1, followed by the results of the characterization study in Sect. 2.2. Trace elements sampled with the described impactor were analyzed with SR-XRF (see Sect. 3.1). For a quantitative analysis, raw spectral count rates have to be linked to ambient concentrations. The production of adequate reference standards for a consistent elemental mass calibration under different experimental conditions is the main focus of this paper, discussed in Sects. 3.2 and 3.3. Mass calibrated time resolved and size-segregated impactor data were finally compared to 24-h filter data in Sect. 4.
2 Sampling method

2.1 Impactor characterization

The rotating drum impactor (RDI, Bukowiecki et al., 2005, 2009) is a modification of the 3-stage UC Davis Rotating Drum Unit for Monitoring (3DRUM Cahill et al., 1985). The 3DRUM was designed for continuous sampling of ambient aerosols in three size ranges of aerodynamic diameter: 2.5–1.15 µm, 1.15–0.34 µm, and 0.34 to approximately 0.1 µm with a sample flow of 22.7 l min⁻¹. The objective of the modification was to design a new impactor with step-wise rotation for a volumetric flow of 16.6 l min⁻¹ (corresponding to 1 m³ h⁻¹) and particle size segregation in ranges of 10–2.5 µm, 2.5–1 µm, and 1 to approximately 0.1 µm. The three impactor stages will be referred to as stage 10 (PM₁₀−₂.₅), stage 2.5 (PM₂.₅−₁) and stage 1 (PM₁−₀.₁) in the following. Particulate matter ≥10 µm is removed by the PM₁₀-inlet (Digitel) on top of the instrument. Due to the stepwise movement of the drum, aerosol particles are deposited in a bar-code-like structure on the film, as illustrated in Fig. 1. Modified RDI drums were designed to be used for sampling as well as for subsequent SR-XRF analysis, see Bukowiecki et al. (2008) and Sect. 3.1. These notched aluminum wheels allow the beam to pass through the wheels without interaction with metallic wheel material. They are covered with a 6-µm polypropylene (PP) film coated with Apiezon M (M & I Materials Ltd.), a silicon-free hydrocarbon grease, to reduce sampling losses due to bouncing effects. One wheel has a capacity for 96 sample bars. An advantage of the combination of RDI sampling on customized wheels with subsequent SR-XRF analysis is that measurements can take place without further sample treatment, unlike most conventional techniques such as inductively coupled plasma optical emission spectrometry (ICP-OES) and mass spectrometry (ICP-MS), reducing the risks of contamination and loss of analyte.

The impaction parameter, or Stokes number, is defined for an impactor as the ratio of the particle stopping distance at the average nozzle exit velocity \( U \) and the nozzle
half-width, $D_j/2$:

$$\text{Stk}_{50} = \frac{\tau U}{D_j/2} = \frac{\rho_p d_{50}^2 U C_c(d_{50})}{9 \eta D_j}$$

(1)

with $\tau$ being the relaxation time, $\rho_p$ the particle density, $d_{50}$ the cutoff diameter, $C_c$ the Cunningham-slip correction factor and $\eta$ the viscosity of air. Under the assumption that the Stokes number remains the same for a given impactor design, the nozzle dimensions for the new RDI were derived through Eq. (1): first, Stokes numbers were calculated for the setup of the original 3DRUM. Based on this, the dimensions for the rectangular nozzles of the RDI with new cutoff sizes were determined accordingly as $(1.52 \times 10)$ mm, $(0.68 \times 10)$ mm and $(0.3 \times 10)$ mm, see Table 1 for details. Since the cutoff diameter $d_{50}$ of the lowest stage of the 3DRUM is not known precisely, it was estimated to lie in the range between 0.06 and 0.12 $\mu$m. Due to this uncertainty, the $d_{50}$ of the RDI for the lowest stage is expected to lie between 0.1 and 0.2 $\mu$m.

Hinds (1982) suggested an ideal Stokes number of $\text{Stk}_{50}=0.59$ for 50% collection efficiency for impactors with rectangular nozzles. To obtain a similar Stokes number for stage 1 as well, either a pressure of approximately 10 kPa and an unrealistically high jet-velocity would be necessary, which could provoke significant particle bouncing. Alternatively, the nozzle size would have to be reduced to $\leq 0.1$ mm for an operation of the impactor at realistic pressure conditions (around 80 kPa). Small-sized nozzles were not practical because of the risk of blockage when the instrument was employed in longer field campaigns. Moreover, the use of multiple small-area nozzles per impactor stage to achieve a more ideal Stokes number under low pressure conditions (as e.g. used in the ELPI, Marjamäki et al., 2000) was not favorable, as the small deposition areas would add additional uncertainty to the applied analysis which produces the best results for large and homogeneous deposition areas. Thus, the low Stokes number for stage 1 (0.06–0.14) is a result of the constraints of the coupled sampling and analysis method considered here at the expense of a reduced impaction efficiency.
2.2 Determination of cutoff sizes

RDI characterization studies were previously conducted using laboratory room air as a quasi-stable proxy for urban ambient air (Bukowiecki et al., 2009). This experimental approach was suitable for the scope of that study, but the necessary corrections through the use of room air restrict a general application. For this study, the cutoff diameters for stages 10 and 2.5 were determined through application of a condensation monodisperse aerosol generator (CMAG, TSI Inc., Model 3475), as in Kwon et al. (2003), and an aerodynamic particle sizer (APS, TSI Inc., Model 3321). Particles produced by the CMAG were quasi-monodisperse dioctyl sebacate (DEHS, C_{26}H_{50}O_{4}) droplets in the size range from approx. 0.3 to 5 μm (geometric standard deviation $\sigma_g=1.4$). Settings for the CMAG varied within the following values: saturator flow 2.25–3 l min$^{-1}$, saturator temperature 235–240 °C, while the reheater temperature remained constant at 100 °C. Average particle concentration produced by the CMAG was 400 particles cm$^{-3}$ after a dilution stage avoiding a too high concentration in the APS. For the cutoff determination of stage 1, particles in the order of 0.1 μm were required, which are difficult to produce with the CMAG. For this purpose, polydisperse NaCl particles were produced with a nebulizer, directed through a dryer and measured with a scanning mobility particle sizer (SMPS) consisting of a differential mobility analyzer (DMA, TSI Inc., Model 3081) and a condensation particle counter (CPC, TSI Inc., Model 3025, high flow). However, higher particle bouncing is expected for NaCl particles because they do not stick as well to the substrate as the oil droplets.

While the APS is a suitable device to measure coarse particles (APS size interval: 0.542–19.8 μm, aerodynamic diameter), the SMPS is a more adequate choice to sample particles in the fine fraction (size interval of employed SMPS: 7–300 nm, mobility diameter). The mobility diameter $d_{\text{mob}}$ measured with the SMPS was transformed into
an aerodynamic diameter $d_p$ using the following recursive equation:

$$d_p = d_{mob} \sqrt{C_c(d_{mob})/C_c(d_p)} \sqrt{\rho_p/\rho_0}$$  

(2)

taking into account the respective Cunningham-slip-correction factors $C_c(d_p)$ and $C_c(d_{mob})$, the density of air $\rho_0$ and the density of DEHS ($\rho_p = 0.91 \text{g cm}^{-3}$) and NaCl particles ($\rho_p = 2.16 \text{g cm}^{-3}$), respectively.

Figure 2 displays the schematic experimental setup, which is essentially the same as in Bukowiecki et al. (2009). For the average particle concentration (reference measurement) the APS measured directly after the CMAG (subsequently referred to as setup A). In setup B, the APS measured the particle size distribution in a sequential manner after stages 10 or 2.5 and then the SMPS measured after stage 1. Switching the APS and SMPS back and forth from setup A to setup B eliminated differences previously encountered by using two different APS instruments (both APS Model 3321). Measurements in each setup were repeated at least 6 times, where a single measurement lasted for about 600 s. Time intervals were 5 s for each APS sample and the SMPS scanning interval lasted about 300 s.

Instruments were connected with conductive tubing (TSI Inc.) and attached to specially manufactured RDI stage covers. Experiments were performed at ambient temperature ($\approx 25 ^\circ \text{C}$). The flow was measured with a primary flow calibrator (A.P. Buck Inc.) and regulated with a mass flow controller (red-y, Vögtlin Ltd.) before each experiment to assure the necessary flow and pressure conditions for a correct operation of the RDI (i.e. $16.6 \text{l min}^{-1}$). This is especially important for stage 1 because of the low cutoff diameter (0.1–0.2 µm). Here, the jet velocity is much higher (106 m s$^{-1}$ compared to 18 m s$^{-1}$, stage 10, and 42 m s$^{-1}$, stage 2.5) and the pressure drops from about 101 to 88 kPa. Stage 1 is close to the transition from an impaction to a diffusion controlled deposition regime. To ensure that identical pressure and flow conditions are maintained at the respective nozzles in both setups, inlet flow rate and pressure were monitored and adjusted accordingly. In order to accommodate for the additional flow
rates when the particle sizers are connected to setup B (5 and 11 min\(^{-1}\) for the APS and SMPS, respectively), the RDI was connected also in setup A.

Size separation characteristics were obtained from collection efficiency curves plotted versus aerodynamic diameter. The cutoff diameter and the stage penetration midpoint diameter \(d_{50}\) both imply the diameter where 50% of particles are collected and 50% pass through (Hinds, 1982). Inverse efficiency curves \(E_i(d_p)\) were computed based on the ratio of the averaged particle size counts measured in setup B and setup A:

\[
E_i(d_p) = \frac{\text{counts}_{\text{setup B},i}}{\text{counts}_{\text{setup A}}}
\]

with \(i\) being either stages 10, 2.5 or 1. Following the concept introduced previously by Bukowiecki et al. (2009), \(E_i(d_p)\) is also referred to as stage penetration at a given particle size \(d_p\), i.e. the non-deposited particle fraction. Kwon et al. (2003) suggested a sigmoidal fit for the inverse efficiency curve:

\[
E_i(d_p) = \frac{a_i}{1 + \exp\left(-\frac{(d_p - c_i)}{b_i}\right)}
\]

with parameters \(a_i, b_i, c_i\), tuned to provide the best fit to the experimental data. In this approach the \(d_{50,i}\) are obtained directly by the parameter \(c_i\), the inflection point. This is equivalent to calculating the inflection point as zero point of the second derivative \(E''_i(d_p)\). All calculated cutoff diameters are compiled in Table 1. Stage penetration curves \(E_i(d_p)\) and second derivatives \(E''_i(d_p)\) of the sigmoidal fits are displayed in Figs. 3 and 4. Already with little noise in the data pronounced peaks occur in the derivative, thus the concept is only applicable for the parts of the curve with very little noise. The fit in the region of diameters smaller than 100 nm for stage 1 is not as satisfactory as for the other stages because variations in the size distribution in the Aitken mode (particles < 100 nm) can have considerable effects on the ratio.
If all particles larger than the cutoff diameter were impacted and all smaller particles passed through, the impactor would have perfectly sharp size cuts and a step-function efficiency curve. Deviations from this ideal theoretical step-function result in a number of oversized particles that pass through and a number of undersized particles that are collected, quantities which ideally should be symmetric. Bukowiecki et al. (2009) introduced the first derivative of $E_i(d_p)$ as a measure for the cutoff sharpness, which is plotted in Fig. 5 versus the normalized diameter $d_p/d_{50,i}$. For better readability all first derivatives $E_i'(d_p)$ were normalized to 1. The curves suggest that the cutoff sharpness achieved with the tested RDI is rather broad, especially for stage 1, which was already observed by Bukowiecki et al. (2009). An implication is that the size limits of individual RDI stages might be slightly smeared out and that smaller particles could be deposited on a higher stage and vice-versa. Several comparisons to concentrations of 24-h filters obtained with high-volume samplers for PM$_{10}$, PM$_{2.5}$ and PM$_{1}$ showed that this effect is not significant on a mass base, see also Sect. 4 and Bukowiecki et al. (2005). The values derived for the cutoff diameters of stage 10 (2.4±0.2 µm) and stage 2.5 (1.03±0.02 µm) correspond well with the theoretical values 2.5 and 1.0 µm, respectively. They confirm the values obtained in the previous study: 2.4±0.2 µm and 1.0±0.02 µm. For stage 1 a value of 0.20±0.02 µm was found, which lies within the expected range of 0.1–0.2 µm.

Through the use of two aerosol generators, the respective aerosol concentrations were high enough to obtain cutoff diameters directly from the data without any further corrections. The presented results verify a correct size segregation within the RDI.

3 Technique for mass calibration

3.1 Synchrotron radiation based X-ray fluorescence spectrometry (SR-XRF)

The low aerosol mass on individual RDI bars demands a highly sensitive detection method. Additionally, a method for a high-throughput analysis was required for RDI
sampling in field campaigns, where the typical number of individual samples can easily reach about 5000. SR-XRF provides a sufficiently high sensitivity and easy sample handling, but it depends on many parameters and therefore requires external calibration in case not all of them are known (Rousseau et al., 1996). Calibration through model calculations of mass absorption coefficients, excitation factors and instrumental characteristics (fundamental parameter analysis) was not applicable due to the high variability in the elemental composition of sampled particles. Today adequate reference materials in terms of similar elemental composition, particle size, sample homogeneity and substrate thickness for ambient aerosol analysis on PP films are still scarce. Thus a general, reusable reference, similar to the sample in terms of matrix composition and sample thickness for calibration of each experimental setup and session was developed to obtain a correct quantification. This provides comparability between different analyses since conditions during experimental sessions at different synchrotron radiation facilities, such as excitation energy, photon flux and detector efficiency can change significantly. This section explains the basic aspects of the applied SR-XRF setup, followed by a description of the development of a technique for external and internal reference element quantification (Sects. 3.2 and 3.3).

Kα lines of a wide range of elements (Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Ni, Cu, Zn, Ga, Sr, Zr, Mo, Cd, Sn, Sb and Ba) as well as the Lα line of Pb were detected with XRF at two synchrotron facilities. Lighter elements (Al-Cr) were measured with a silicon drift detector (Roentec Xflash 2001 type 1102, Bruker AXS) with a nominal resolution of 155 eV (Mn Kα at 5.9 keV) at the optics beamline in the Swiss Light Source (SLS) at the Paul Scherrer Institute, Switzerland. Total available photon energy ranges from 5.5 to 22.5 keV at a bending magnet with a cryogenically cooled Si(111) channel cut monochromator (Flechsig et al., 2009). Trace elements were examined in the focused monochromatic mode, with an energy of 11.5 keV and usable photon flux of $2\times10^{11}$ photons s$^{-1}$ within a 70×140 μm ($h\times v$, FWHM) focus area. A helium atmosphere was applied to reduce absorption effects in air and to eliminate the Ar fluorescence line at 2.9 keV (occurring with measurement in air). Heavier elements (Mn-Pb)
were measured at the beamline L, HASYLAB, DORIS III storage ring at DESY where a bending magnet provides a polychromatic spectrum with usable photon statistics up to 80 keV. Al and Cu absorbers of various thicknesses (high-pass filter) can be used to shift the energy maximum towards higher energies and thus reducing background effects from lower energies. For measurements of the data set presented in Sect. 4 an 8-mm Al absorber was used. A polychromatic 100×200 µm (h×v) wide beam irradiated the sample and a nitrogen cooled Si(Li)-detector (Sirius 80, Gresham) with a nominal resolution of 144 eV (Mn Kα) measured the fluorescence counts. This detector is suitable to measure Kα lines of heavier elements (up to the Ba Kα line at 32 keV) due to the large active volume (Si crystal depth: 4 mm, area: 80 mm²) and a Be window thickness of 12.5 µm. Since the detection efficiency of Si(Li) detectors decreases for lighter elements as electronic noise increases, a Si drift detector with a 3-µm polymer window (plus a 0.5-µm aluminum layer) and smaller active volume (depth: 450 µm, area: 10 mm²) with the ability to process higher count rates was employed for measurements at the SLS. Low peaking times (in the order of \( \tau_p = 1 \mu s \)) are suitable for this detector, but would lead to a too low energy resolution in the Si(Li) detector, which was operated with \( \tau_p = 12 \mu s \). The efficiency of Si drift detectors decreases rapidly for higher energies because the thin Si crystal absorbs less than 30% above 11 keV. Thus, each detector is matched to the chosen excitation energies. The detector deadtime was kept below 30% by reducing the size of the photon beam. Output and input count rates (OCR, ICR) were measured while varying the opening of exit slits to investigate the detector saturation regime. Determining the relationship between OCR and ICR for a given detector at given experimental conditions enabled correcting for potential dead time effects in the detector.

Sample wheels were rotated with a goniometer in steps of 3.51°, corresponding to the separation of individual RDI bars, and each bar was irradiated typically for 20–30 s. Since synchrotron radiation is linearly polarized, positioning the fluorescence detector in the polarization plane at an angle of 90° with respect to the incident beam reduced the spectral background due to coherent and incoherent scattering substantially.
3.2 Absolute mass calibration

Fluorescence counts of sample elements can be linked to the area concentration (µg cm\(^{-2}\)) if the deposited mass of one sample, chosen as reference, is determined externally by wet-chemical methods in a subsequent step. An example is ICP-OES, but it requires more analyte mass than deposited on a single RDI bar, because the sensitivity is not high enough for these minute masses. Furthermore, the sample material digestion and possible contamination error demanded a new, non-destructive method without structural modification. Since readily available calibration films of similar composition and thickness as the sample do not exist, producing a customized calibration film became necessary. Fittschen et al. (2006, 2010) introduced a concept for applying picoliter droplets via an ink-jet printer on different reflector substrates for TXRF. Transforming this approach to the substrates used in SR-XRF by Bukowiecki et al. (2008) gave rise to a procedure of applying standard solutions on thin films with an ink-jet printer. Calibration films were fixed on RDI sample wheels to safeguard that experimental conditions for the SR-XRF analysis are the same.

The use of a Compact Disc-label printer (HP Photosmart D5160), along with the film structure itself ensured optimized adherence of the solution on the substrate because the film is not bended and transported by small brushes as in conventional printers. Clean printer cartridges (type HP 339, completely cleaned by Pelikan Ltd.) with a 15-pl ink drop volume were filled with customized solutions. A precondition for a correct calibration is the similarity of reference and aerosol specimen in terms of homogeneity (grain size and shape), chemical composition and concentration. This would imply using standards with a limited concentration range adapted to the concentration range of the samples. In contrast, a decreasing uncertainty of slope and intercept is obtained for more measurement points in a wider concentration range (Van Grieken, 2002). Therefore, the concentration range was chosen as a compromise between similarity to the sample and reduction of the uncertainty. In this range the relation between fluorescence intensity and mass per printed area is expected to behave linearly.
Two techniques to obtain several increasing coating densities of standard solutions on the substrate were tested: printing one to five times the same amount of solution or printing five areas with different transparency (i.e. “color” saturation in the printer settings). Only the first technique yielded the requested linear increase in mass per coated area. For each of the five coatings pieces of 9 cm$^2$ from the same printing process were analyzed by ICP-OES. The linear relationship of fluorescence intensity (normalized to photon flux and detector dead time) for the reference element Cd versus the obtained mass per analyzed area ranging from 0.11 to 0.33 µg cm$^{-2}$ is plotted in Fig. 6. The slope of the fitted linear curve of reference element counts ($\Phi_{st}$) versus mass per area in µg cm$^{-2}$ ($C_{st}$) is inserted into the calibration formula introduced below (Eq. 7).

When primary X-rays interact with material, scattering and secondary fluorescence excitation occurs beside primary absorption (Giauque et al., 1979). This implies that the fluorescence intensity of the aerosol sample is not only based on the absolute concentration but can also depend on the total chemical composition (i.e. the matrix seen as sample plus substrate). Owing to the thin sample layer and low mass density of collected particles, the samples are irradiated completely and concentrations are too low to cause secondary X-ray absorption by lighter elements (neglecting the effects of particle size). In the investigated ambient trace element determination, matrix effects were neglected as discussed already by Bukowiecki et al. (2008).

Scattering leads to low peak to background ratios. To keep this scattering as low as possible, it is advisable to use the thinnest substrate possible. However, the thinner the substrate, the more difficult the handling of the film is, so printing inhomogeneities can occur. The chosen substrates represent a compromise between a thin film and satisfactory printing results. Commercially available films in different thickness were tested: a 100-µm ink jet transparency film (3M, CG3420), a 100-µm PET ink jet film (Folex, BG-32.5 RS plus), both coated, a 50-µm PET film (Folex, X-131) and a 25-µm self adherent polypropylene film, both uncoated. Some films contain interfering elements like Si (adhesive of the self adherent film), S, Al, and Ca, which are also found
Extensive tests of different substrates, solutions and printing processes led to three applicable reference samples, which will be discussed in the following. First, a multi-element solution (Merck standard IV, containing Ag, Al, B, Ba, Bi, Ca, Cd, Co, Cr, Cu, Fe, Ga, In, K, Li, Mg, Mn, Na, Ni, Pb, Sr, Ti, Zn plus single elements: P, Rb, S, Sb, Se, Sn, Ti, Zr) was applied on the 100-µm PET film with a high-resolution printing process (1200 dpi). Next, a self adherent 25-µm PP film was tested. Again, the Merck standard solution IV was the basis and Rb and Se plasma-standards were added (10 g l⁻¹ in HNO₃ solution) with equal concentration for each element. Through addition of 0.5% of Triton X-100, a tenside to decrease surface tension, a more homogenous wetting and improved drying speed was achieved. The buffered solution had a pH of 2 as a compromise between the risk of corrosion of cartridges caused by a low pH and elemental precipitation/separation caused by a high pH. Count rates of both reference materials (on 100- and 25-µm films) are illustrated in Fig. 7, indicating a much more articulated response curve for the thinner PP film due to reduced scattering. Also, the high Sb peak caused by the 100-µm PET substrate vanished for the thinner film. For measurements of the 100-µm film the beam exit slit size had to be reduced significantly to avoid saturation of the detector due to high count rates. These count rates were extrapolated according to the determined non-linear relationship between slit size and count rates due to deadtime effects for the further analysis. The chosen standard solution led to satisfactory results for the calibration of heavier elements. However, to calibrate the lighter elements more precisely, a solution containing fewer elements was prepared, avoiding the interference of Lα lines with the Kα lines in focus. Again, clean printer cartridges were filled with a customized solution containing...
K, Ca and Ti standards (10 g l\(^{-1}\) in HNO\(_3\) solution), Si (1 g l\(^{-1}\) in HNO\(_3\)/HF solution) and Al (1 g l\(^{-1}\) in HNO\(_3\) solution). Building on the good results obtained before, this solution was applied on the self adherent 25-µm film. The resulting calibration curve is discussed in the next section (Figs. 8 and 9). This reference led to a satisfactory result for lighter elements and showed the high potential of customized calibration solutions for specialized purposes.

3.3 Relative calibration based on external standardization

In addition to the absolute mass calibration, a relative calibration is necessary because the fluorescence yield increases with increasing atomic number Z. The main reason is the Auger effect, a process competing with fluorescence in which the photon is absorbed within the atom and the released energy is emitted through an Auger electron. Absorption of the photon within the atom is most pronounced for lighter atoms, and significantly limits the yield of secondary X-rays from the lighter elements. The fluorescence yield \(\omega\) is the relative frequency of photon emission (in competition with the relative frequency of Auger electron emission \(\chi\), Bambynek et al., 1972; Burhop et al., 1955). An approximation of \(\omega\) is given by:

\[
\omega = \frac{Z^4}{A + Z^4}
\]  
(5)

with a constant \(A=9\times10^5\) for the K-series and \(A=7\times10^7\) for the L-series. Since Auger electron and fluorescence photon emission are two complementary processes, \(\omega + \chi = 1\). Because the calibration films (with similar substrate thickness and matrix compared to the samples) contain a series of elements with the same concentration, it is possible to experimentally determine the response curve and calibrate the count rates with a relative factor \((S_{rel})\). The empirically determined \(S_{rel}\) comprises all influences on the effective fluorescence intensity such as theoretical fluorescence yields, mass attenuation caused by sample elements and detector sensitivity. The absorbed photon
fraction $P_{\text{abs}}$ was calculated based on the source parameters of the DORIS III storage ring plus the 8-mm Al absorber inserted into the beam path (Sánchez del Río et al., 2004) and mass attenuation coefficients of X-rays for chosen elements in the sample (Berger et al., 2007):

$$P_{\text{abs}} = \sum_{E_i < E(\Phi_0)} \left( \Phi_0 x_s - \frac{\Phi_0}{\mu \rho} \left( 1 - e^{-\mu \rho x_s} \right) \right)$$

(6)

In this approximation for infinite thin films, the contributions of absorption caused by the relevant polychromatic photon intensities $\Phi_0$ for typical elements were summed up. Other variables in the equation are the sample thickness $x_s$, the total absorption mass attenuation coefficient $\mu$ and the material density $\rho$. The calculated absorbed photon fraction reflects the overall trend of the empirical curve.

All three calculated curves were added to Fig. 8 showing the empirical relative calibration curve of Kα lines for measurements at HASYLAB. A sigmoidal or exponential fit to the data points turned out to provide the best extrapolation of the relative curve for those sample elements that are not contained in the calibration solution. The energy where the detector efficiency drops to zero ($E_0$) is included as the lower limit of the fit. Since Cd was chosen as reference element in the absolute mass calibration, the count rates for $S_{\text{rel}}$ in Fig. 8 are normalized to the count rate of Cd.

Figure 9 displays the relative calibration curve $S_{\text{rel}}$ obtained by analyzing the calibration reference customized for lighter elements with SR-XRF at the SLS. Here, a similar curve to previous experiments was found and through the fit to data, all elements in the sample can be analyzed.

Coating homogeneity, reproducibility of printed areas and stability of printed surface are challenges in the printing process. Scanning electron microscope (SEM) images from printed calibration and blank films revealed a sufficiently good homogeneity of applied droplets. The presented results show that customized calibration films for different experimental SR-XRF conditions can be produced with the ink-jet printer method.
3.4 Calibration formula

Spectra were fitted with the WinAxil software package (Canberra Inc., Van Espen et al., 1986; Vekemans et al., 2004) using a least squares fitting algorithm with a Bremsstrahlung background. Although the Bremsstrahlung background originally has its source in the description of electron induced X-rays, where retardation of the electrons is almost completely responsible for the continuum, this model was able to reconstruct the background curvature best (optical inspection). Spectral counts obtained by WinAxil are calibrated with the absolute mass calibration factor $C_{st}/\Phi_{st}$ and the relative calibration factor $S_{rel}$. The ambient concentration $C$ of one element is deduced from the fluorescence intensity $\Phi$ by the following calibration formula:

$$C = \frac{\Phi C_{st} A_i t_c}{\Phi_{st} S_{rel} A_c t_m} \left( \frac{1}{t_{RDI} Q_{RDI}} \right) \left( \frac{100}{100 - t_d} \right) \left( \frac{l_m}{l_D} \right)$$

(7)

with the RDI bar area ($A_{10} = 15.2 \text{ mm}^2$, $A_{2.5} = 6.8 \text{ mm}^2$ and $A_1 = 3.0 \text{ mm}^2$), the total calibration film area analyzed by ICP-OES ($A_c = 9 \text{ cm}^2$), the respective irradiation times for aerosol ($t_m$) and calibration spectra ($t_c$), the RDI sampling interval ($t_{RDI}$), the RDI flow ($Q_{RDI}$), the dead time caused in the detector ($t_d$), the actual beam current ($I_D$) and the maximum beam current directly after injection ($I_m$). The last term $l_m/l_D$ is only applied to measurements performed at HASYLAB, where the raw count rates have to be normalized to the photon flux. No correction is necessary for measurements at SLS because of a constant beam current due to top-up injection. As mentioned before this calibration technique is only applicable for references with similar elemental matrix and similar film thickness.

Measurement uncertainties were calculated with uncertainty propagation of the three terms in Eq. (7) containing an uncertainty. The extrapolation from the rather small-sized beam spot to the RDI bar area ($A_{\text{beam}} \approx 1\%$ of total area $A_i$) adds to the total measurement uncertainty with a contribution of 10% of the area. Further uncertainty is introduced by possible slight variations in the inlet flow, which is estimated to contribute...
a relative uncertainty of 5%. The third term is the uncertainty obtained by the linear regression for the calculation of the absolute mass calibration factor.

Minimal detection limits (MDL) were determined as a means for the qualitative evaluation of every individual data point. Only elements exceeding the MDL with >50% of values remained in the final data set. These detection limits are calculated as follows:

\[
\text{MDL} = \frac{C_{st} A_i}{\Phi_{st} S_{rel} A_c t_{RDI} Q_{RDI}} 3 \sqrt{\Phi_B} \quad (8)
\]

with \(\Phi_B\) being the elemental continuum counts obtained by fitting a Bremsstrahlung background. Thus, a longer counting time improves the MDL for a fixed setup.

### 4 Results and discussion

As shown in Sect. 2.2 the RDI segregates the sampled ambient aerosols in three size ranges. This size segregation enables a more detailed interpretation of the relative elemental composition. Quantitative ambient aerosol measurements can be obtained through application of the calibration for the SR-XRF method described in Sect. 3.

To consolidate these two findings, exemplary results are presented. The RDI was deployed in a field campaign at Zürich Kaserne, Switzerland for time and size resolved trace element sampling from 28 November 2008 to 5 January 2009 (with a short break from 26 to 28 December) and a time resolution of 2 h.

The relative elemental composition for the three size ranges is illustrated in Fig. 10. Note that for these pie charts, the period from 31 December 2008 15:00 LT to 1 January 2009 05:00 LT is excluded to not distort the picture through unusually high emissions of some elements (e.g. S, K, Ti, Cu, Sr, Ba) during the fireworks at New Year’s Eve. Nonetheless, sulphur and potassium account for the highest contributions in the fine size range. Secondary sulfate and biomass combustion emissions are assumed to contribute primarily to the high sulphur concentration. Ammonium sulfate is formed by conversion of \(\text{SO}_2\) to sulfate via either heterogeneous reactions in droplets (with ozone,
NO$_2$, H$_2$O$_2$) or photochemically via OH radicals (Seinfeld, 1998). Fine potassium also originates mainly from combustion processes. In the coarse size range, Cl (presumably from de-icing salt) and Fe contribute the highest amounts to the mass. This gain in information through the size resolution of the RDI enhances the potential of source apportionment studies significantly (Han et al., 2005; Ondov et al., 2006; Karanasiou et al., 2009) and will be exploited in an upcoming study.

In order to validate the obtained concentrations, results from RDI measurements were compared to independent 24-h filter data from the same campaign. Two high-volume samplers (Digitel, Aerosol Sampler DHA-80) with PM$_1$ and PM$_{10}$ inlets were used to collect 24-h samples on 1, 3, 5, 7, 9, 11, 13, 15 and 17 December 2008. A fraction of the quartz micro-fiber filter was acid digested and subsequently analyzed by ICP-OES and ICP-MS (see details of the method in Querol et al., 2008) for the determination of major and minor elements. Only PM$_{10}$ data for typical elements (S, K, Fe, Cu, Sn and Sb) are shown because the number of data points above the detection limit is limited for the PM$_1$ filter data. RDI data were binned into 24-h intervals by calculating the mean of 12 data points, and all three RDI stages were summed up for the comparison to PM$_{10}$. Time series for the chosen elements are visualized in Fig. 11. Uncertainty bars in Fig. 11 are calculated as described above for RDI data, while the uncertainty for filter data was calculated following the method of Escrig et al. (2009). Two outlier days (7 and 9 December) show a higher deviation than others which may be due to increased sample inhomogeneity or other unknown experimental issues. Correlations of RDI data versus filter values show no clear trend towards a general over- or under-estimation by the RDI analysis. The gain in information obtained by a higher time resolution, which can lead to the identification of diurnal variations, is a major advantage of the RDI method. The overall comparison of 24-h concentration values shows a reasonable agreement of both methods and therefore confirms the applicability of the presented calibration methodology.
5 Conclusions

The cutoff diameters of a rotating drum impactor were measured through quasi monodisperse DEHS particles from a condensation monodisperse aerosol generator and polydisperse NaCl particles from a nebulizer with an aerodynamic particle sizer and a scanning mobility particle sizer to consolidate previous studies carried out in laboratory room air. Results show a good agreement with theoretical values and confirm the validity of the earlier experiment (Bukowiecki et al., 2009). However, the cutoff sharpness for the analyzed impactor is broader than that found for other cascade impactors.

The developed calibration method for SR-XRF count rates to ambient concentrations properly matched the requirements and provided an external, reusable reference material for the presented studies. The aim of obtaining consistent data sets from different experimental sessions for subsequent source apportionment studies was reached. The procedure is advantageous over already existing standards as the elemental composition and the thickness of the substrate is comparable to the sample composition and the film used for sampling. However, when using a calibration film that does not contain sufficient elements in the region of interest, adaptation to additional data, such as filter values may be necessary. Printing standard solutions on a simple transparency film with a conventional Compact Disk-label printer forms an easy to handle approach for both relative and absolute mass calibration. Averaging a series of spectra leads to a satisfactory result, although the printing homogeneity might be improved in the future by using a professional large-scale printer. While the substrate film thickness was appropriate in the presented case, a professional printer with the possibility to fix the film through electrostatic adhesion or a vacuum sample holder could lead to the use of a thinner film where lower background effects are expected.

The comparison of high time resolved and size-segregated RDI data to 24-h filter data shows a good agreement. Despite the need for access to a synchrotron facility to conduct SR-XRF analysis for high time resolution trace element data, the results
demonstrate that the gain in information compared to conventional low time resolution wet-chemical evaluation justifies these efforts. High time resolution data enables a better identification of sources because daily patterns such as peak-traffic hours and other day-night differences in anthropogenic activities can be observed. The data will be used to perform source apportionment with positive matrix factorization (PMF) in an upcoming study.

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**References**


Giauque, R., Garrett, R. B., and Goda, L. Y.: Determination of trace elements in light element matrices by X-ray fluorescence spectrometry with incoherent scattered radiation as an inter-
References


Table 1. Values of characteristic parameters of 3DRUM and RDI as well as penetration midpoint diameters in µm (aerodynamic diameter) calculated from the second derivative $E''(d_p)$ of the sigmoidal fit for RDI stages 10, 2.5, and 1. The free parameter $c_i$ of the sigmoidal fit, which is an equivalent way to obtain the inflection point, is listed as well.

<table>
<thead>
<tr>
<th></th>
<th>stage 10</th>
<th>stage 2.5</th>
<th>stage 1</th>
</tr>
</thead>
<tbody>
<tr>
<td>3DRUM $d_{50}$ [µm]</td>
<td>1.1</td>
<td>0.34</td>
<td>0.06–0.12</td>
</tr>
<tr>
<td>3DRUM $D_j$ [mm]</td>
<td>0.82</td>
<td>0.31</td>
<td>0.25</td>
</tr>
<tr>
<td>Stk$_{50}$(3DRUM,RDI)</td>
<td>0.49</td>
<td>0.44</td>
<td>0.06–0.14</td>
</tr>
<tr>
<td>RDI $d_{50}$ [µm]</td>
<td>2.5</td>
<td>1.0</td>
<td>0.1–0.2</td>
</tr>
<tr>
<td>RDI $D_j$ [mm]</td>
<td>1.52</td>
<td>0.68</td>
<td>0.3</td>
</tr>
<tr>
<td>$d_p$ from $E''(d_p)$ fit</td>
<td>2.4±0.2</td>
<td>1.03±0.02</td>
<td>0.20±0.02</td>
</tr>
<tr>
<td>$d_p$ from $c_i$</td>
<td>2.4±0.2</td>
<td>1.03±0.02</td>
<td>0.20±0.02</td>
</tr>
</tbody>
</table>
Fig. 1. RDI sampling drum: a notched aluminum wheel, coated with a 6-µm PP film, used for sampling and subsequent SR-XRF-analysis as well as for the calibration. The bar code-like structure of deposited particulate matter is visible on the film. The black color on the depicted stage 1 bars with a width of 300 µm and a height of 10 mm is mainly caused by soot deposition.
**Fig. 2.** Schematic layout of the two subsequent setups for cutoff determination experiments: the left side illustrates the reference measurement (setup A), where the concentration of oil droplets produced by the CMAG is measured directly with the APS. For stage 1 the NaCl concentration produced by a nebulizer is measured with the SMPS. The RDI is connected in order to simulate realistic flow conditions. The right side displays the actual experimental setup, where the APS is connected subsequently to the covers of stages 10 and 2.5 and then the SMPS is connected to the cover of stage 1 (setup B).
Fig. 3. Stage penetration curve $E_i(d_p)$ for stages 10 and 2.5: ratio of size distributions obtained in setup B versus size distributions obtained in setup A. The upper panel (a) shows the ratio for stage 10, a sigmoidal fit was applied to data points. The second derivative $E''_i(d_p)$ is shown on the same axis scale. The lower panel (b) shows data points and sigmoidal fit for stage 2.5, $E''_{2.5}(d_p)$ is shown on the same axis scale.
Fig. 4. Stage penetration curve $E_1(d_p)$ for stage 1: ratio of size distributions obtained with the SMPS connected to the RDI (setup B) versus size distributions obtained with the SMPS connected to the nebulizer (setup A). The upper panel (a) shows the data points and the sigmoidal fit. For better readability the second derivative $E''_1(d_p)$ is displayed in the lower panel (b).
Fig. 5. Experimental RDI cutoff sharpness: normalized first derivatives of $E_i'(d_p)$ ($E_i'(d_p)_{\text{max}} = 1$) for all RDI stages 10, 2.5 and 1 plotted versus the normalized diameter $d_p/d_{50,i}$. The $E_i'(d_p)$ curves represent a measure for the size-segregation sharpness in the RDI.
Fig. 6. Fluorescence intensity of reference element Cd: individual data points represent the average count rates (normalized to photon flux and detector dead time) for about 50 spectra of each coating on the calibration film. The absolute mass calibration factor is obtained by the slope of the linear regression (with an intercept of zero) of mean fluorescence count rates in cps versus area concentration in µg cm$^{-2}$. 
Fig. 7. Comparison of SR-XRF spectra of two different calibration films: count rate of elements in the Merck IV standard solution on a 100-µm PET film and the count rate of essentially the same solution on a 25-µm PP film. Count rates measured at HASYLAB are normalized to photon flux and detector dead time.
Fig. 8. Relative calibration curve $S_{rel}$ of Ka lines for measurements at HASYLAB: count rates relative to the count rates of the reference element Cd are plotted versus atomic number Z with the corresponding fit. Theoretical curves for the fluorescence yield as calculated with Eq. (5), absorbed photon fraction as calculated with Eq. (6) (estimation for experimental conditions taking into account the HASYLAB excitation spectrum with a 8-mm Al absorber) and detector efficiency are shown on the same axis scale.
Fig. 9. Relative calibration curve $S_{rel}$ of Kα lines for measurements at SLS: count rates are normalized to detector dead time and plotted relative to the count rate of Ti, blank values are always subtracted. An exponential fit was applied to the data and shows the extrapolated relative calibration curve $S_{rel}$ for all elements in the sample.
Fig. 10. Pie charts for relative contributions of measured elements in the three size ranges ($\text{PM}_{10-2.5}$, $\text{PM}_{2.5-1}$ and $\text{PM}_{1-0.1}$) from the winter campaign in Zürich Kaserne obtained by RDI-SR-XRF trace element analysis. The percentage of measured trace elements from the total PM$_{10}$ mass lies in the order of 5–10 %.
Fig. 11. Comparison of time series of PM$_{10}$ filter data and the sum of all RDI stages from Zürich winter campaign for typical elements. Each data point corresponds to one day, i.e. one filter value and the mean of 12 RDI values.